



# The influence of thermal extremes on coral reef fish behaviour in the Arabian/Persian Gulf

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**Abstract** Despite increasing environmental variability within marine ecosystems, little is known about how coral reef fish species will cope with future climate scenarios. The Arabian/Persian Gulf is an extreme environment, providing an opportunity to study fish behaviour on reefs with seasonal temperature ranges which include both values above the mortality threshold of Indo-Pacific reef fish, and values below the optimum temperature for growth. Summer temperatures in the Gulf are comparable to those predicted for the tropical ocean by 2090–2099. Using field observations in winter, spring and summer, and laboratory

experiments, we examined the foraging activity, distance from refugia and resting time of *Pomacentrus trichrourus* (pale-tail damselfish). Observations of fish behaviour in natural conditions showed that individuals substantially reduced distance from refugia and feeding rate and increased resting time at sub-optimal environmental temperatures in winter (average SST = 21 °C) and summer (average SST = 34 °C), while showing high movement and feeding activity in spring (average SST = 27 °C). Diet was dominated by plankton in winter and spring, while fish used both plankton and benthic trophic resources in summer. These findings were corroborated under laboratory conditions: in a replicated aquarium experiment, time away from refugia and activity were significantly higher at 28 °C (i.e. spring temperature conditions) compared to 21 °C (i.e. winter temperature conditions). Our findings suggest that *P. trichrourus* may have adapted to the Arabian/Persian Gulf environment by downregulating costly activity during winter and summer and upregulating activity and increasing energy stores in spring. Such adaptive behavioural plasticity may be an important factor in the persistence of populations within increasing environmentally variable coral reef ecosystems.

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## Introduction

Over the last century, sustained climate change has resulted in average global sea surface temperature increasing by 0.6 °C, with predictions of further 4.0 °C rises by 2100 (Collins et al. 2013; IPCC 2014). As the planet warms, climate and weather variability, as well as the occurrence

of extreme weather events, will also increase (IPCC 2012; Thornton et al. 2014). As the world's oceans absorb 93% of excess heat trapped by greenhouse gases (Laffoley and Baxter 2016), such warming is expected to have major consequences for the biodiversity, functional structure and productivity of marine ecosystems (Folguera et al. 2011; Doney et al. 2012; Bellard et al. 2012; Pecl et al. 2017). Alterations have already been documented in the diversity and assemblage structure in temperate seaweed and seagrass ecosystems (Wernberg et al. 2013; Vergés et al. 2014), coral reefs (Hughes et al. 2018) and coastal temperate rocky reefs (Lejeune et al. 2010; Albouy et al. 2015). Despite this, our understanding of the capacity for marine ecosystems to resist or show resilience to climate change is limited (Côté and Darling 2010).

Coral reefs are particularly susceptible to a range of natural and anthropogenic disturbances that have resulted in recent degradation and loss of coral on a global scale (Wilkinson 1999; Cheal et al. 2017; Hughes et al. 2018). While these disturbances are clearly significant for the corals that serve as the biogenic engineers of these ecosystems, diverse and ecologically important reef-associated fish assemblages also often exhibit dramatic changes in composition and abundance with increasing environmental variation (Day et al. 2018; Richardson et al. 2018; Gordon et al. 2018). Indeed, in addition to susceptibility to habitat loss (Pratchett et al. 2018), coral reef fishes are also directly threatened by ocean warming, sea surface temperature anomalies and intense pulse heat stress events (such as the warm phase of El Niño Southern Oscillation—ENSO) (Claar et al. 2018; Cai et al. 2018), due to their narrow thermal tolerance range (Tewksbury et al. 2008; Sunday et al. 2011). Experimental evidence suggests that coral reef fish may already be living close to their upper thermal limits (Rummer et al. 2014) and that their population structure may be directly impacted by increased variability in water temperature (Folguera et al. 2011; Pratchett et al. 2015; Rodgers et al. 2018).

One of the primary responses to environmental temperature change is modification of behaviour, with the speed and scope of behavioural adjustment potentially determining survival (Tuomainen and Candolin 2011; Wong and Candolin 2015). Despite this, tropical reef fishes having evolved in relatively thermally stable environments are expected to show little ability to adjust behaviour adaptively with changing environmental temperatures (Candolin 2018). Furthermore, as the metabolism and neurophysiology of fishes are directly influenced by temperature (Pörtner and Farrell 2008; Pörtner et al. 2010), environmental temperatures outside an individual's optimum range may then have a substantially negative effect on ecological performance and survival via non-adaptive (dysfunctional) modification of behaviour, resulting in

altered feeding rate, increased risk-taking behaviour and reduced activity levels and habitat use (Figueira et al. 2009; Nagelkerken and Munday 2016). Despite this, there is a growing body of research suggesting that coral reef fishes may show adaptive behavioural plasticity in response to temperature increases, predominantly associated with temporary reductions in aerobically costly behaviours with acute increased environmental variance (Johansen et al. 2014; Scott et al. 2017; Chase et al. 2018). For example, feeding is associated with substantial metabolic costs due to the high energy expenditure needed for digestion and nutrient uptake (McLean et al. 2018) and has been shown to be downregulated in some fish species to reduce energy costs (Nowicki et al. 2012; Fu et al. 2018). Nevertheless, how well and by what mechanisms coral reef fishes will cope with ocean warming scenarios and increased thermal variability (Thornton et al. 2014; IPCC 2014; Cai et al. 2018) is still poorly understood.

One approach to understanding how individuals may behaviourally respond to environmental variance is to study contemporary communities that exist within naturally variable environments. The fauna associated with the high-latitude reefs of the southern Arabian/Persian Gulf (hereafter the 'Gulf') provide a natural test of the impact of extreme water temperature variation on coral reef fish behaviour (Feary et al. 2010; Burt et al. 2011b). The Gulf experiences the highest annual change in water temperature for coral reefs globally (winter: < 15 °C, summer: > 35 °C), with fishes persisting for several months in conditions that would be considered lethal to reef fishes in other parts of the world (Riegl and Purkis 2012; Vaughan et al. 2019). Summer temperatures in the Gulf exceed the upper thermal limits of most tropical reef fishes (Nilsson et al. 2009; Munday et al. 2009; Rummer et al. 2014; Rodgers et al. 2018), while winter temperatures are well below the optimum temperatures reported for reef fish elsewhere in the Indo-Pacific (Eme and Bennett 2008; Figueira and Booth 2010; Nakamura et al. 2013). Indeed, contemporary summer water temperatures in the Gulf are comparable to those predicted for tropical oceans at the end of this century (Riegl and Purkis 2012; IPCC 2014), while winter conditions can be so severe as to induce cold water coral bleaching (Shinn 1976; Coles and Fadlallah 1991).

Using a combination of in situ field observations and ex situ experimental manipulations, this study examined the influence of extreme thermal variability on reef fish behaviour in the southern Gulf. Using the locally abundant *Pomacentrus trichrourus* (pale-tail damselfish; Günther, 1867), we performed field-based observations of behaviour across three thermally distinct seasons separated by a 13 °C sea surface temperature (SST) range (winter: SST = 21 °C, spring: 27 °C, and summer: 34 °C), monitoring replicate individual's distance from refugia, resting

times, feeding rate and diet between three sites in the southern Gulf. We predicted that fish would respond to low and high water temperature extremes by modifying their behaviour (i.e. exhibit behavioural plasticity). Specifically, we hypothesised that fish would minimise aerobically costly behaviours (i.e. reduced distance from refugia and feeding rate, while increasing resting time) when exposed to sub-optimal temperatures for growth and locomotion in winter and summer, while compensating for energy losses and building up energetic stores by increasing activity and feeding rate in spring. We also hypothesised that *P. trichrourus* would exhibit flexibility in diet by shifting from a predominately planktonic diet during optimal conditions to one encompassing both planktonic and benthic prey, when exposed to sub-optimal temperature. Such change in diet was expected to be due to restrictions in foraging movements in sub-optimal conditions, within individuals becoming generalist in their feeding habits, exploiting more easily accessible resources. Finally, we performed an aquarium-based experiment exposing *P. trichrourus* to temperatures typical of winter (21 °C) and spring (28 °C) in order to examine the hypothesis that variation in behaviour observed in the field was directly mediated by temperature. Overall, the results of this study have important implications for understanding how coral reef fishes may use behavioural modification as a means to respond to the environmental variability that is expected to increase under future climate change.

## Materials and methods

The pale-tail damselfish is a small omnivorous damselfish common in southern Gulf coral communities, often living in association with a single ‘home’ coral head (Allen 1991; Shraim et al. 2017). Within the Gulf, this species shows high abundance, a benthic lifestyle with diurnal behaviour, a strong association with the reef, and high territoriality to a small area (territory size of 0.5–1.5 m<sup>2</sup>). It has a small body size (max 6 cm standard length [SL]) and distinct differences in colour pattern between juvenile and adults (Randall 1995; Feary et al. 2010). All individuals within this study were adults (~ 4–6 cm SL). The optimum temperature for growth for this species is modelled to be between 24.7 and 29.3 °C (mean 27.2 °C) (Kaschner et al. 2016, based on 258 cells from the Western Indian Ocean, including the Red Sea and the Gulf).

### In situ behavioural quantification of *P. trichrourus*

To determine the effect of seasonal changes in temperature on behaviour of *P. trichrourus* in the Gulf, behavioural observations were replicated across three seasons: spring

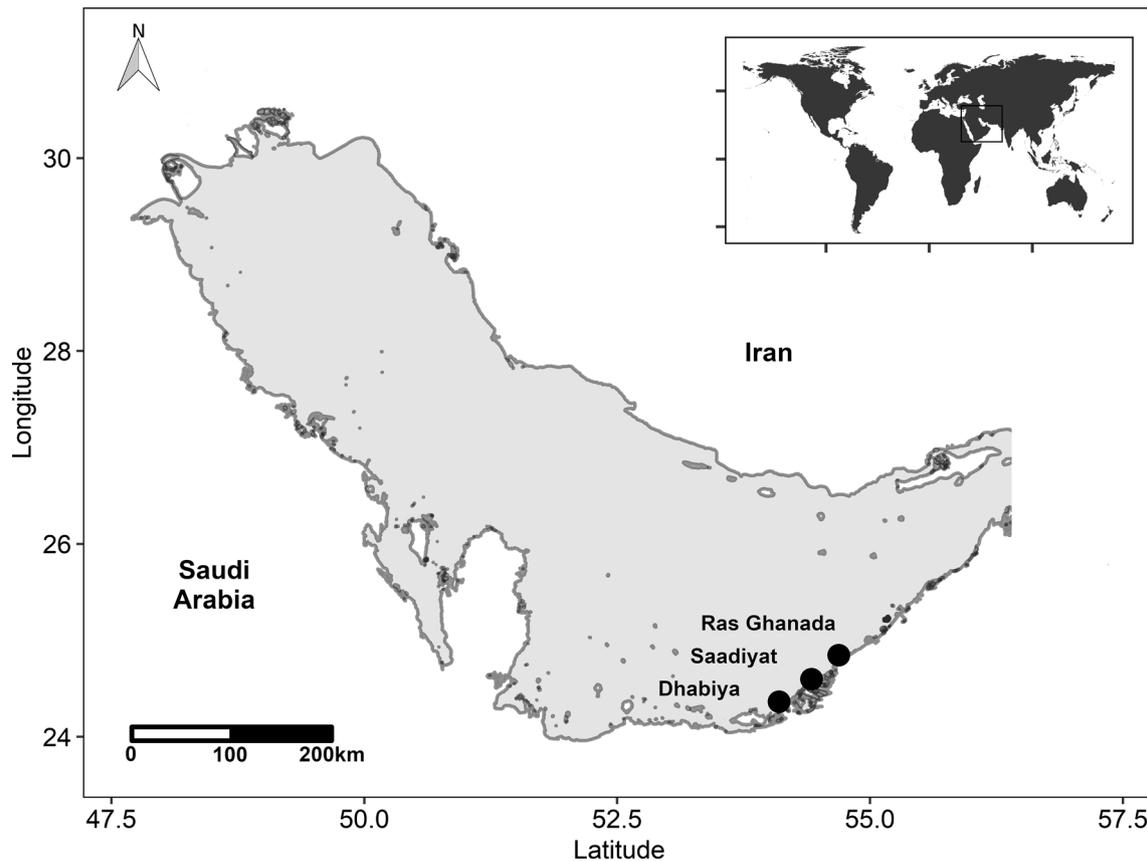
(April/May 2016, average SST = 27 °C), summer (August/September 2016, average SST = 34 °C) and winter (February/March 2017, average SST = 21 °C). Observations were undertaken at three reef sites spaced ca. 40 km apart along the Abu Dhabi coastline (United Arab Emirates—southern Gulf): Dhabiya (24°21′55.8″N, 54°06′02.9″E), Saadiyat (24°35′56.4″N, 54°25′17.4″E) and Ras Ghanada (24°50′53.4″N, 54°41′25.1″E) (Fig. 1). Sites were located at similar distances from shore (between 5.8 and 3.8 km), with available habitat dominated by hard-coral communities (ca. 40–55% live coral cover, Burt et al. 2011a) with all sites 6–7 m in depth (Grizzle et al. 2015).

### In situ behavioural observation

Observations were undertaken by scuba diving on cloudless days between 10:00 and 14:00, reducing any influence of changing light levels on behaviour (Layton and Fulton 2014). In each site, across seasons, replicate coral heads holding *P. trichrourus* individuals were identified by visual searches made while swimming along haphazardly located transects of the reef. Coral heads held one (56% of coral heads examined), two (34%), three (7%) or four (2%) adult *P. trichrourus*. Individual *P. trichrourus* were haphazardly selected during each season and their territory identified and marked using numbered flags (Magalhaes et al. 2013). When several adults shared the same coral head, in order to minimise the developmental variability among individuals being sampled, the individual with the largest body size was selected. Following marking of the territory, focal fish was allowed to acclimate for 4 min, and then filmed for 2 min (using DBPOWER HD 1080P). A minimum distance of 3 m was maintained between observer and subject. Observations were aborted if fish displayed any adverse reactions to diver’s presence (i.e. increased wariness, hiding after staring at diver, etc.; observed in < 5% of individuals). Across sites and seasons, a total of 226 individual behavioural observations were recorded (winter: Dhabiya *n* = 26, Saadiyat *n* = 32, Ras Ghanada *n* = 25; spring: Dhabiya *n* = 26, Saadiyat *n* = 32, Ras Ghanada *n* = 25; and summer: Dhabiya *n* = 23, Saadiyat *n* = 31, Ras Ghanada *n* = 29).

### Analyses of video quantifying in situ individual behaviour

Mean distance from refugia was defined as the distance from the ‘home’ coral every 10 s (Beck et al. 2016) and estimated to the nearest 5 cm referenced against the territory marking flag’s length (25 cm) and the individual’s body size (assumed as 5 cm SL). Resting was taken as total time (seconds) spent immobile (i.e. not active swimming, foraging or chasing behaviour) across each 2-min video. Feeding rate was the total number of bites during the



**Fig. 1** Study area and sampling locations in the southern part of the Persian Gulf

observation period, while diet was identified by classifying each food bite as either on pelagic or benthic substrates. Benthic food bites were further classified as being on live coral, coral rubble, turf or sand. Behavioural data were initially extracted from 15% of videos and independently analysed and validated by two observers. As results differed by  $< 5\%$  between observers, all remaining video analyses were carried out by one observer (DD).

To estimate the relative frequency of predation or competition across seasons and sites, the abundance of predators (i.e. piscivorous fish with  $SL \geq 15$  cm observed interacting with *P. trichourus*) and competitors (i.e. fishes with comparable ecology/body size to *P. trichourus* observed chasing/being chased by *P. trichourus*) entering each individual territory was quantified for each 2-min video observation.

#### *In situ environmental measurements*

To assess seasonal changes in abiotic variables in the Gulf, across the dates of all behavioural observations, daily in situ SST ( $^{\circ}\text{C}$ ) and dissolved oxygen (DO,  $\text{mg l}^{-1}$ ) measurements were taken at 1200 h at each site using a YSI Professional Plus multi-parameter probe, while

underwater visibility (in m) was visually estimated in situ by swimming over a belt transect and measuring the distance at which the diver waiting at the beginning of the transect was not visible anymore. Chlorophyll-a (chl-a,  $\text{mg l}^{-1}$ ) measurements were downloaded as monthly composite data at 0.05 resolution from the MODIS-Aqua satellite (NASA 2014), while salinity (psu), wind ( $\text{m s}^{-1}$ ) and currents ( $\text{m s}^{-1}$ ) were obtained from the NCEP climate forecast system version 2 (CFSv2, Saha et al. 2011). Salinity and current data were accessible as daily means (resolution  $0.25\text{-deg} \times 0.25\text{-deg}$ ), while wind data were taken at 1200 h ( $0.205\text{-deg} \times \sim 0.204\text{-deg}$ ) across all experimental days.

#### **Aquarium-based behavioural quantification**

To experimentally determine the impact of water temperature on behaviour, we examined time away from refugia and distance moved of replicate *P. trichourus* in controlled conditions in the laboratory. Sixteen adults (mean body size  $4.78 [\pm 0.13 \text{ SE}]$  cm SL) were collected from Saadiyat during winter. Eight individuals were randomly allocated to separate pierced 6L bottles within either of two 900L holding tanks (control and experimental group, i.e.

one holding tank per treatment). Each bottle had a 10-cm-long plastic pipe provided for shelter. All tanks were kept at 21 °C, 40 psu, 8.15 pH, 6.5 mg l<sup>-1</sup> DO and 12 h dark/light cycle reflecting in situ winter condition. 25% water changes were performed weekly, with ammonia, nitrite and nitrate kept at 0 ppt. Fish were fed ad libitum twice daily with commercial pellets.

After a week for acclimation, refuge use and movement of all individuals were tested at 21 °C (Fig. 2). To do this, a randomly chosen individual had its holding container submerged in an experimental tank (80 L \* 50 W \* 40 H [cm], filled to 20 cm H) and was left to acclimate for 10 min. The tank's base had been divided into 40 squares (10 \* 10 cm grid); a 10-cm-long segment of plastic pipe was provided as shelter, and a DBPOWER HD 1080P video camera was placed one metre above the tank. Following acclimation, the individual was released and led to the shelter; a transparent bottomless bottle was used to confine the individual which was then left for a further 20 min. Subsequently, the bottle was removed and behaviour videoed for 16 min. All videoing occurred between 1000 h and 1400 h, with a 25% water change between individuals.

Water temperature within the experimental holding tank was then increased to 28 °C (~ 0.5 °C/6 h) across 4 d, while the temperature within the control remained constant (21 °C) (Fig. 2). Fish were then left for 7 d. As described above, all individuals were then retested for behaviours within either the experimental or control tank, using individual videoing.

Within all videos, the first 5 min was discarded to allow time of acclimation after activation of the video camera. Time away from refuge was quantified (in seconds) when individuals moved 10 cm from the shelter. Distance moved was quantified (in cm) by counting the number of grid lines crossed over 11 min with one cross being assumed to equal

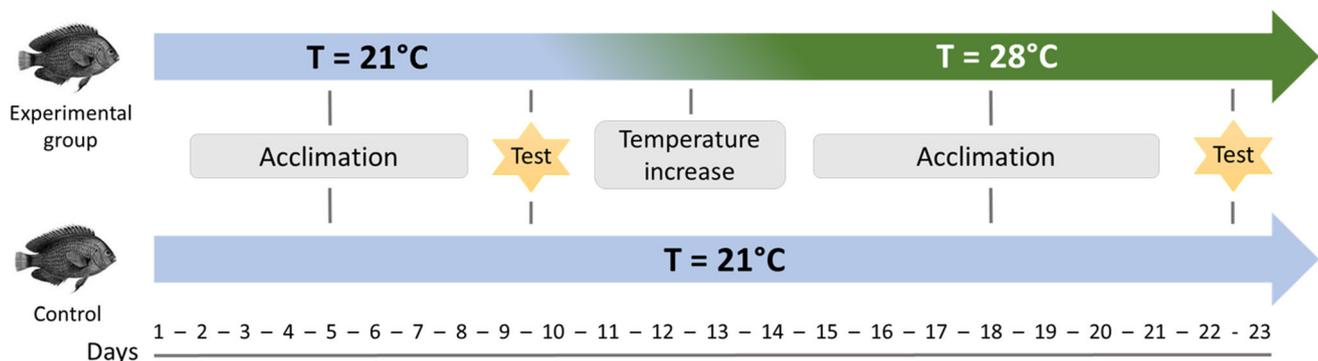
to 10 cm movement, independently of the distance from refuge.

### Statistical analysis

To determine whether abiotic and/or biotic factors significantly differed among seasons, Kruskal–Wallis test was undertaken. Abiotic factors were then combined using principal components analysis (based on the correlation matrix with centring and scaling of all variables), to identify the main axes of environmental variation among observations; axes that explained more than 15% of the total variance (PC1 to PC3, hereafter ‘environmental PCs’) were retained and used in further analyses (Magalhaes et al. 2016).

To determine the impact of season, site, predator/competitor abundance and environmental PCs on distance from refugia and resting time, a generalised linear mixed model (GLMM) with Gaussian errors was used, while feeding rate against all predictor variables (as above) was examined using a GLMM with Poisson error distribution and log link function. Both distance from refugia and resting were not normally distributed and were, respectively, log and log (X + 1) transformed prior to testing. In all tests, date of sampling was fitted as a random effect and backwards model selection followed, using likelihood ratio tests to examine the significance of each term removed from the model. To determine whether food preference significantly differed between seasons, individual bites on substrate were analysed using Kruskal–Wallis test.

To determine the effect of experimental change in temperature on time away from refugia and distance moved, behavioural data before and after temperature manipulation from the control and experimental groups were tested using a GLMM with binomial error distribution and logit link function. As time away from refuge and distance moved were bimodal with a high proportion of



**Fig. 2** Experimental design of the laboratory behavioural experiment. Days (bottom) represent the timeline of the experiment; bottom arrow represents the control group; middle boxes represent the

different activities to which the fish in the two different groups were exposed; top arrow represents the experimental group

zeros, and data transformation did not result in either normality or homoscedasticity, data were converted to binary; an individual was considered ‘far from refugia’ or ‘moving’ when time away from refugia or distance moved (respectively) was more than the median and vice versa. In all tests, ‘test order’ was included as a fixed factor to determine whether the order in which individual fish were tested had any influence on the result, while individual was fitted as a random effect; backwards model selection followed, using likelihood ratio tests to examine the significance of each term removed from the model. All data were analysed in R (R core development team 3.5.1, 2019).

## Results

### In situ environmental variation

Significant differences among seasons in SST, DO, chl-a, salinity and visibility were apparent (Kruskal–Wallis test:  $p < 0.05$  in all cases; Table S1). SST was highest in summer and lowest in winter, while DO showed the opposite trend. Chl-a concentration was higher in summer than the other seasons. PCA showed that 82% of the variation in the raw abiotic variables was captured by the first three principal components, with a high score in PC1 (39%) being associated with low SST and chl-a, and high DO. A high score in PC2 (25%) was associated with high salinity and low visibility, and a high score in PC3 (18%) was associated with high water current (Table 1).

Across all sites and seasons, predator abundance was dominated by *Lutjanus ehrenbergii*, comprising 97% of counts (Table S2). Predator abundance significantly differed among seasons (Kruskal–Wallis test,  $H = 73.790$ ,  $df = 2$ ,  $p < 0.001$ ), with predators being more abundant in summer than spring or winter, while among sites, predators were more abundant in Dhabiya than Saadiyat or Ras Ghanada (Fig. S1). The predominant competitors were

other *P. trichrourus* and *Pomacentrus aquilis* individuals. Competitor abundance significantly differs among seasons ( $H = 7.191$ ,  $df = 2$ ,  $p = 0.027$ ), with slightly higher abundance in spring than winter (Table S2).

### In situ behavioural analysis

Season, and the interaction between site and season, had a significant effect on *P. trichrourus* behaviour (Table 2), with individual from Ras Ghanada and Saadiyat showing higher distance from refuge and higher feeding rate in spring than in either winter or summer, while significantly reducing the time spent resting (Fig. 3). On the contrary, individuals from Dhabiya showed similar behavioural patterns across seasons in feeding rate only, but not in distance from refuge nor time spent resting (Fig. 3).

Significant differences in diet were apparent among seasons, with planktonic prey representing 93% and 96% of food bites in winter and spring, respectively, while only accounting for 73% in summer (Kruskal–Wallis,  $H = 139.04$ ,  $df = 2$ ,  $p < 0.001$ ). Where selected, benthic food bites were predominantly taken from live corals (5 and 4% of total food bites in winter and spring, 17% in summer), with bites on coral rubble, turf and sand accounting for less than 1% of total food bites in winter and spring, and  $\leq 4\%$  in summer.

### Laboratory behavioural assays

While there was no difference between the control and the experimental group prior to temperature manipulation (i.e. both groups exposed to cold temperature), temperature had a significant effect on fish behaviour after manipulation, with fish spending more time far from the refuge and moving longer distances in the warm experimental treatment than the control (Fig. 4a, b; Table 3). Test order had no effect on fish behaviour.

## Discussion

Although increasing variance in global SSTs may directly impact coral reef fish community biodiversity (Pratchett et al. 2015; Day et al. 2018; Buchanan et al. 2019), coral reef fish populations might be able to mitigate the effect of rapidly changing environments through behavioural plasticity (Nagelkerken and Munday 2016; Scott et al. 2017; Chase et al. 2018). We examined the impact of seasonal changes in biotic and abiotic conditions on distance from refugia, activity and feeding ecology of the pale-tail damselfish *P. trichrourus* within the southern Gulf. We found that *P. trichrourus* substantially increased resting time and reduced distance from refugia and feeding rate at sub-

**Table 1** Loadings of each environmental variable on the first three PC axes and the per cent variance explained by each axis

Factors	PC1	PC2	PC3
Temperature	– <b>0.58</b>	– 0.06	– 0.04
Dissolved oxygen	<b>0.58</b>	0.11	0.06
Chl-a	– <b>0.49</b>	0.25	0.03
Salinity	– 0.10	<b>0.70</b>	0.14
Wind speed	– 0.25	– 0.22	– 0.26
Water current	0.06	– 0.14	– <b>0.83</b>
Visibility	– 0.08	– <b>0.60</b>	<b>0.47</b>

The highest loadings on each of the first three PCs are showed in bold

**Table 2** Results of GLMM analysis of effects of biotic/abiotic factors on *Pomacentrus trichrourus*' in situ behaviour

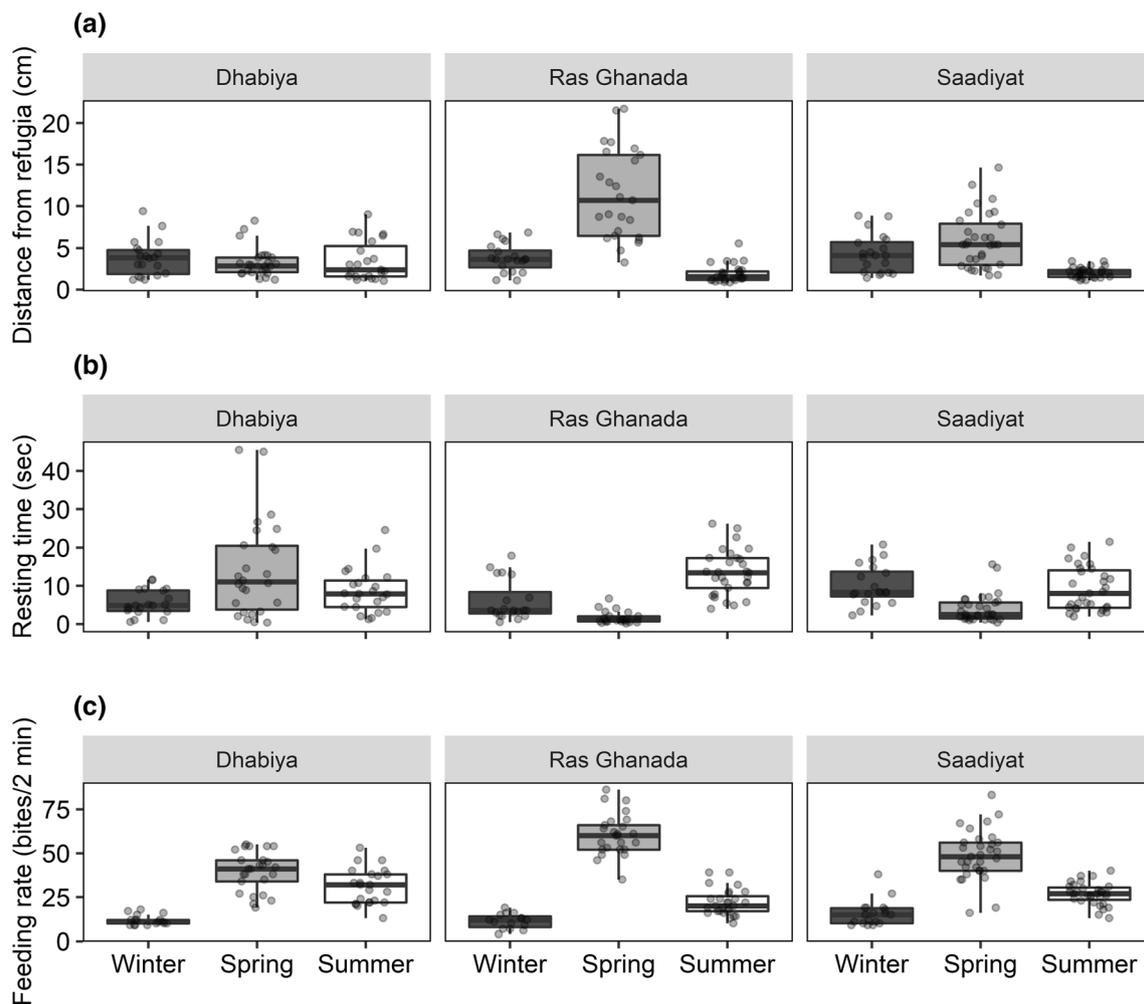
Response	Predictor	Residual deviance <sub>(df)</sub>	Change in deviance <sub>(df)</sub>	<i>p</i>
Distance from refugia	Season	8.298 <sub>(5)</sub>	10.925 <sub>(2)</sub>	<b>0.004</b>
	Site	− 1.1050 <sub>(5)</sub>	1.522 <sub>(2)</sub>	0.467
	Site: season	− 2.627 <sub>(7)</sub>	30.565 <sub>(4)</sub>	< <b>0.001</b>
	PC1	− 37.767 <sub>(14)</sub>	0.430 <sub>(1)</sub>	0.512
	PC2	− 33.192 <sub>(11)</sub>	3.690 <sub>(1)</sub>	0.055
	PC3	− 37.390 <sub>(13)</sub>	0.377 <sub>(1)</sub>	0.539
	Predators	− 36.882 <sub>(12)</sub>	0.508 <sub>(1)</sub>	0.476
	Competitors	− 38.197 <sub>(15)</sub>	0.176 <sub>(1)</sub>	0.674
Resting time	Season	89.580 <sub>(5)</sub>	4.384 <sub>(2)</sub>	0.112
	Site	86.431 <sub>(5)</sub>	0.539 <sub>(6)</sub>	0.539
	Site: season	85.196 <sub>(7)</sub>	29.556 <sub>(4)</sub>	< <b>0.001</b>
	PC1	54.937 <sub>(12)</sub>	1.013 <sub>(1)</sub>	0.314
	PC2	55.640 <sub>(11)</sub>	0.703 <sub>(1)</sub>	0.408
	PC3	53.924 <sub>(13)</sub>	1.324 <sub>(1)</sub>	0.25
	Predators	52.042 <sub>(15)</sub>	0.015 <sub>(1)</sub>	0.908
	Competitors	52.601 <sub>(14)</sub>	0.558 <sub>(1)</sub>	0.455
Feeding rate	Season	1730.2 <sub>(4)</sub>	31.427 <sub>(2)</sub>	< <b>0.001</b>
	Site	1700.7 <sub>(4)</sub>	1.883 <sub>(2)</sub>	0.39
	Site: season	1698.8 <sub>(6)</sub>	17.959 <sub>(4)</sub>	<b>0.001</b>
	PC1	1674.6 <sub>(12)</sub>	1.459 <sub>(1)</sub>	0.227
	PC2	1680.8 <sub>(10)</sub>	3.805 <sub>(1)</sub>	0.051
	PC3	1677.0 <sub>(11)</sub>	2.487 <sub>(1)</sub>	0.115
	Predators	1672.9 <sub>(14)</sub>	0.137 <sub>(1)</sub>	0.137
	Competitors	1673.1 <sub>(13)</sub>	0.165 <sub>(1)</sub>	0.685

Gaussian error structure for distance from refugia and resting time; Poisson error distribution with log link function for feeding rate. Significant *p* values are highlighted in bold

optimal environmental temperatures (both high and low) when observed in situ and when tested in the laboratory. Individuals were also observed to shift from a planktonic-based diet to a mixed planktonic and benthic diet during the Gulf's extreme summer season. This reflects the finding of a recent stomach content study in this species which showed seasonal shifts in diets in the southern Gulf, with more benthic species (e.g. coral) in the diet in summer, while more planktonic species (e.g. free swimming ascidians, various protozoans) were more common in cooler months (Shraim et al. 2017). Our results show considerable scope for behavioural and trophic plasticity in *P. trichrourus*, which may allow this species to cope with seasonal variance in environmental conditions, including those that are above the mortality thresholds (Nilsson et al. 2009; Munday et al. 2009; Rummer et al. 2014; Rodgers et al. 2018) and below the optimum temperature of Indo-Pacific reef fish (Eme and Bennett 2008; Figueira and Booth 2010; Nakamura et al. 2013).

We observed substantial behavioural and trophic plasticity in *P. trichrourus*' populations among seasons. In particular, *P. trichrourus* were observed to reduce distance from refugia and feeding rate and increase resting time in

winter and summer compared to spring. These seasonal changes are plausibly related to variation in water temperature, as changes in environmental temperature are known to have a direct effect on fish physiology (Angilletta 2009; Abram et al. 2017). In ectothermic animals, environmental temperatures departing from the optimum are associated with reductions in the capacity to deliver oxygen to aerobic processes and, therefore, reduction in energy for costly activities (i.e. swimming, feeding, chasing, etc.) (Pörtner and Knust 2007; Pörtner and Farrell 2008; Munday et al. 2012). The reduction in activity observed in *P. trichrourus*, is also in line with recent work suggesting that, when exposed to short-term higher than average SSTs, coral reef fish may temporarily mitigate bioenergetic inefficiency through reduction in energetically costly activities, as well as through changes in diet (Nowicki et al. 2012; Johansen et al. 2014; Scott et al. 2017; Chase et al. 2018). However, further research is now needed to confirm whether *P. trichrourus* actively reduce energetically costly behaviour to minimise energetic demand and mitigate bioenergetic inefficacy, or whether fish are being forced to change their behaviour due to being compromised at a lower level of organization (i.e. physiological constraints).



**Fig. 3** Changes in *Pomacentrus trichrourus* behaviour in situ among seasons (winter = dark grey, spring = light grey, summer = white) and locations (Dhabiya, Ras Ghanada, Saadiyat): **a** mean distance from refugia (cm); **b** time spent resting (s/120 s) and **c** number of

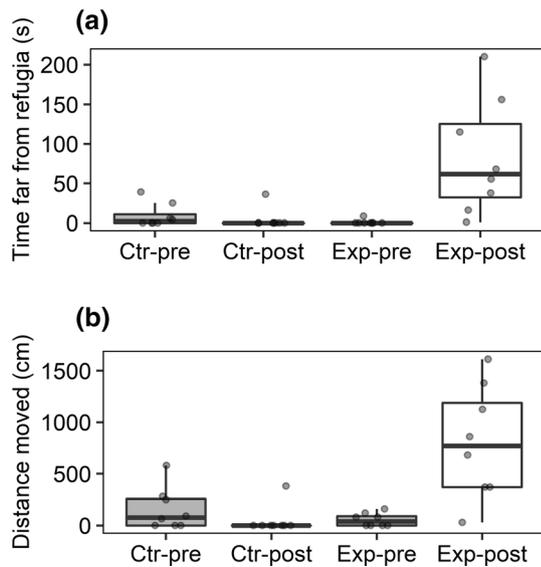
feeding bites in 120 s. Thick horizontal lines show the median, and boxes show interquartile (IQR). Whiskers indicate the range of data, and dots show outliers that are more than 1.5 IQR above the 75th percentile, or more than 1.5 IQR below the 25th percentile

Nevertheless, even if fish are able to modulate their activity to reduce the degree of stress they are experiencing, such strategies may not be sufficient to sustain prolonged periods (i.e. months) of extreme temperature (Rummer et al. 2014; Rodgers et al. 2018). For example, reductions in food intake and quality, coupled with low aerobic scope under extreme temperatures, can disrupt an ectothermic organism's energy balance, with substantially less energy available for long-term maintenance, growth and reproduction (Gillooly et al. 2001; Pörtner et al. 2010; Neuheimer et al. 2011), impacting individual fitness and potentially survival (Chase et al. 2018).

*Pomacentrus trichrourus*' persistence within the Gulf implies that they have adapted over the long term to the extreme temperature conditions experienced there (Rummer and Munday 2017). Our results hint that the key to their survival may be their ability to undertake seasonal energy recovery in spring, compensating for energy loss

during winter by building up energetic stores to endure the upcoming summer season (Armstrong and Bond 2013). Many studies on temperate fishes have highlighted the ability to compensate energetic and growth losses during periods of adverse temperatures or food scarcity through increased feeding rate and feeding activity when conditions become favourable again (Sevgili et al. 2013; Armstrong and Bond 2013; Furey et al. 2016; Peng et al. 2017). For *P. trichrourus*, such periods of recovery, where it can access highly nutritious food resources (Shraim et al. 2017) under optimal environmental temperatures (Kaschner et al. 2016), may be vital for mitigating energy loss during winter and summer and hence ultimately permit the persistence of abundant populations within the southern Gulf throughout the year.

Our laboratory experiments are consistent with the idea that the seasonal differences in behaviour observed in the field were the result of changes in temperature. However,



**Fig. 4** Changes in *Pomacentrus trichrourus* behaviour ex situ among temperature treatment: **a** mean time spent far from refugia (s) and **b** total distance moved (cm). Colours indicate temperature (grey = 21 °C, white = 28 °C). See Fig. 3 legend for explanation of the box plots

other potentially important environmental variables also changed with the seasons, including predators and primary productivity. These variables could plausibly be directly responsible for changes in fish behaviour. For example, predators might have had a negative effect on *P. trichrourus* movement (i.e. reduced distance from refugia) and feeding rate (Dill and Fraser 1984; Beck et al. 2016; Catano et al. 2016), potentially leading to the shift to mixed planktonic and benthic diets observed in summer, when *P. trichrourus* may have to rely more on safer benthic food resources (i.e. live corals), despite the loss of nutritional value (Shraim et al. 2017). Similarly, the higher abundance in predatory fish in Dhabiya compared to other sites (Burt and Vaughan, unpublished data) may explain why fish were seldom observed moving long distances from the

refuge, even in spring. On the contrary, we did not observe any correlation between changes in primary productivity and changes in feeding activity, suggesting that shifts in diet may not be related to primary productivity (measured as chlorophyll-a concentration) (Rueda et al. 2015; Zhou et al. 2016) but the result of temperature-related differences in food preference (Shraim et al. 2017). Ultimately, manipulative experiments would be needed to confirm the potential for predators or other drivers to produce similar behavioural changes to those seen in our temperature manipulation.

The Gulf’s average SST is increasing twice as fast as the global average (Al-Rashidi et al. 2009), with predictions of further increases of 0.5–1.4 °C by 2050, associated with hot salt brine discharge derived from local desalination plants (AGEDI 2016; Vaughan et al. 2019). As *P. trichrourus* populations live for five months of the year above their predicted thermal optimum (Jun–Oct, mean SST > 30 °C) (Kaschner et al. 2016), of which two months encompass temperatures above the mortality threshold of low-latitude Indo-Pacific damselfish (July and August, mean SST ≥ 33 °C) (Nilsson et al. 2009; Munday et al. 2009; Rummer et al. 2014; Rodgers et al., 2018), Gulf populations may already be living close to their upper thermal limits; further increases in temperature are expected to pose a serious threat to population persistence. In addition, *P. trichrourus* populations in the Gulf are already considered endangered, associated with a fragmented distribution and with a strong dependence on diminishing coral habitat resources (Buchanan et al. 2016, 2019). Indeed, *P. trichrourus* are highly associated with live coral colonies, utilising them for recruitment, shelter and as trophic resources (Buchanan et al. 2016; Shraim et al. 2017). Therefore, any increases in SST that pose a serious threat to local coral communities, where bleaching events and consequent coral mortality are already common (Riegl et al. 2018; Burt et al. 2019; Paparella et al. 2019), may deprive *P. trichrourus*

**Table 3** Results of GLMM analysis of effects of experimental change in temperature on *Pomacentrus trichrourus*’ behaviour

Temperature manipulation	Response	Predictor	Residual deviance <sub>(df)</sub>	Change in deviance <sub>(df)</sub>	<i>p</i>
Pre-	Time far from refuge	Treatment	17.119 <sub>(3)</sub>	2.7561 <sub>(1)</sub>	0.097
		Test order	11.945 <sub>(8)</sub>	5.1733 <sub>(5)</sub>	0.395
	Distance moved	Treatment	20.088 <sub>(3)</sub>	1.0823 <sub>(1)</sub>	0.298
		Test order	18.56 <sub>(8)</sub>	1.5277 <sub>(5)</sub>	0.910
Post-	Time far from refuge	Treatment	6.2268 <sub>(3)</sub>	17.194 <sub>(1)</sub>	<b>&lt; 0.001</b>
		Test order	4.1307 <sub>(10)</sub>	2.0961 <sub>(7)</sub>	0.954
	Distance moved	Treatment	6.2268 <sub>(3)</sub>	17.194 <sub>(1)</sub>	<b>&lt; 0.001</b>
		Test order	4.1307 <sub>(10)</sub>	2.0961 <sub>(7)</sub>	0.954

Significant *p* values are highlighted in bold

‘Treatment’ denotes whether fish belonged to the control or experimental group

populations of suitable habitat and food resources in the long-term (Keith et al. 2018; Pratchett et al. 2018).

The Gulf's marine community encompasses 241 species of coral-associated bony fish, which are considered a biogeographic subset of the Indian Ocean's fauna that recolonised the region between 6000 and 9000 yrs ago (Riegl and Purkis 2012). The coral reef fish community richness in the Gulf is thought to be lower than that in the Indian Ocean as a consequence of the Gulf's physical extremes, which fewer Indian Ocean species are able to tolerate (Coles 2003; Feary et al. 2010). However, despite a range of work examining the reef fish fauna in the region (Burt et al. 2011b; Buchanan et al. 2019), there is still little understanding of how environmental tolerance is achieved in Gulf-dwelling species. For *P. trichourus*, this study has shown that populations may be able to behaviourally mitigate the high variance in temperature and temperature extremes which are consistent with predictions for the tropical ocean by the end of the century (IPCC 2014). It appears that by reducing movement from shelter and feeding activities within extreme seasons (summer, winter) and showing metabolic compensatory activity behaviour (including feeding) during the environmentally benign spring season, populations of *P. trichourus* are able to thrive in one of the most thermally extreme marine environments. Such adaptive behavioural plasticity is consistent with current theory regarding how animals deal with extremes and may be an important mechanism for how increases in temperature and temperature variability may be mitigated by fishes on Indo-Pacific reefs by the end of the century.

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#### Compliance with ethical standards

**Conflict of interest** On behalf of all authors, the corresponding author states that there is no conflict of interest.

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